



Litter decomposition and nutrient cycling in temperate forest of Kumaun region

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Abstract

Background & Aim: The magnitude of litter fall and its decomposition in a forest ecosystem impart a crucial role in nutrient supply to plant as well as microorganisms. The present study was conducted to evaluate and compare the litter fall production, leaves litter decomposition and nutrient re-translocation among the dominant species i.e. Pine and Oak, of temperate forest of Kumaun region. **Results:** The estimated annual litter fall production was significantly 23% higher in Pine stand than Oak stand. The maximum portion of litter fall was occurred in summer and dry period between April-June. The nutrients concentration was higher in green leaves than leaves litter. Additionally, the nutrients concentration was found to be high in green leaves as well as in leaves litter of Oak species. The decomposition rate of leaves litter was found to high in Oak floor which is inversely proportional to the C/N ratio. The nutrient immobilization for all the nutrients was found to be high in Pine stand. The nutrient re-translocation percentage were in the order of P > K > N > Ca, in both the species that indicates the higher immobilization of phosphorous and potassium in the region. Soil status is also responsible for litter decomposition. The soil MBC, MBN, MBP, basal respiration and enzymes activity were higher in Oak stand which is directly correlated with the decomposition rate. **Conclusions:** Our results suggested that different plant species growing in same climatic condition differed in litter fall, litter decomposition and have different level of nutrients uptake and availability.

Keywords: Litter decomposition; Nutrient cycle; retranslocation; *Pinus roxburgii*; *Quercus leucotricophora*

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1. Introduction

The energy required by various ecosystem processes is mainly controlled and available from photosynthesis. Soil does not have capability to capture energy directly from sun i.e. solar energy because of the lack of photosynthetic organisms and depends on the energy rich bodies mainly produced by plants residues. These plant residues are decomposed by a series of naturally occurring biological processes that involve both microbial flora and fauna. Litter fall from leaves and other part of plants is the basic character of all vegetative environments, the vital source of nutrients and the beginning point of nutrient cycle. It is the main pathway through which organic matter and nutrients returns to the soil surface, where it composes the major portion of nutrients available to soil, incorporation with roots

turnover^{1,2}. The availability of nutrient to the soil reflects the rate of decomposition and microbial activity³. Litter fall along with roots composes a major segment of nutrient cycling between plants and soils, thus internal fluctuations of C, N and P at ecosystem level are constrained by the available litter fall^{4,5}. The nutrients availability in any given soil system is mainly because of the decomposition dynamics of organic matter. The major part of carbon and the energy that enters the food web is detritus based litter that are used by the heterotrophic organisms and act as a nutrient reservoir in the food web. Plant litter production and decomposition is a crucial ecosystem process that defines and governs the plant-soil relationships by regulating the nutrient turnover and the buildup of soil organic matter. In fact, recycling of nutrients through decomposition is the primary source of available nutrients for plants in most unmanaged terrestrial ecosystems⁶. In contrast, the ecosystem existence is based on the decomposed matter. The rate of decomposition and the nutrient releases are highly influenced by the tree species through different properties, such as magnitude of litter produced, litter

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quality and nutrients release as well as climatic conditions and existing microbial communities in the soil system⁷.

Forests worldwide are known to be significantly important habitats acts as a storehouse of biodiversity. These forests are the source of a major part of global terrestrial carbon in which temperate forest ecosystem plays major role in carbon sequestration from increasing atmospheric carbon dioxide, as it covers the major portion of terrestrial land. The productivity and the functioning of a forest are mainly influenced by the pattern and the rate of nutrient cycling. In a forest ecosystem, litter fall and its decomposition act as a clogged system which evidently interrelated with the cycling of nutrients; it regulates the nutrients and energy flow between biological components and soil, fixes energy and governs the overall growth and productivity of forest^{8,9}.

The magnitude of nutrient re-translocation and the nutrient use efficiency in litterfall is the key source of nutrients for plants as well as used as an indicator of soil nutrient status¹⁰. Along the same climatic conditions, different plant species respond differently in their nutrient release pattern. The re-translocation of nutrients and their movement via soil and leaf litter is characterized as one of the important process in nutrient dynamics that are prominently used by plants to conserve nutrients.

Numerous studies have been documented on various features of litter fall in different forest ecosystem all over the world as well as in India. However, the information regarding different aspects of litter fall for temperate forest of India is still limited. The present study is undertaken in Nainital district of Kumaun region located in Uttarakhand. The forest covers a diverse range of plantation¹¹. The main objective of the study is to examine the ability of two dominant plant species of the region in their litter fall creation and litter decomposition under same environmental condition and their impact on nutrients availability and uptake.

2. Methods

2.1 Site description

The study was conducted in the unprotected forest of Nainital district, located in Uttarakhand. The pine (*Pinus roxburgii*) and oak (*Quercus leucotrichophora*) are the two most dominated plant species of the region. The selected area is divided into two stands on the basis of dominant plant species. The stand dominated with pine plantation is surrounded between E 079° 32' 22.9", N 29° 23' 24.4" and at elevation of 1822m. Similarly, oak dominated stand is lie within E 079° 33' 08.3", N 29° 21' 15.9" and at elevation of 1333m. The vegetative structure

of the region has been described by Singh and Singh, 1986¹². The climate is summer monsoon with three distinct season categorized into summer, which is warm and dry (March-May), rainy season is warm and wet (June-September) and the winter is cold and moderately dry (October-February). The minimum temperature ranges from -3 °C to 12 °C during the month of January and the maximum temperature fluctuates from 18 °C to 30 °C in the month of June. The average annual rainfall is 151.9mm (Figure. I). The soil is generally brown in color with 58.22%, 11.33% and 30.45% of sand, slit and clay respectively.

2.2 Litter fall collection

To study litter input in the selected forest stands i.e. Pine and Oak forest, five litter traps were randomly placed in each stands within the established plots of 10X10m². The litter accumulated within the traps was collected at regular interval time period for two consecutive years i.e. Oct, 2012 to Sep, 2014. Litter samples were taken to the laboratory oven dried at 60 °C and then weighed. The green leaves/needle samples of Pine and Oak were also collected at the same time.

2.3 Chemical analysis

The dried leaf litter and the green leaves samples were then processed for chemical analysis. The carbon content was measured by loss of ignition method. The nitrogen content was determined using Kjeldahl method¹³. Carbon content was measured by loss of ignition method. Stannous chloride method¹⁴ followed by hot plate digestion in HNO₃:HClO₃ (3:1) at 180 °C for 6 h was used for phosphorous and potassium concentration.

2.4 Litter decomposition

The litter decomposition rate was estimated using nylon bag technique. A total of 50 litter bags (25 bags for each stand) were prepared. Each Litter bag of 20X20cm (5mm mesh) was filled with 10gm of air dried weighed of litter sample and placed on the forest floor in October, 2012. The litter bags were drawn at fixed interval of time during the study period. After collection, the extra

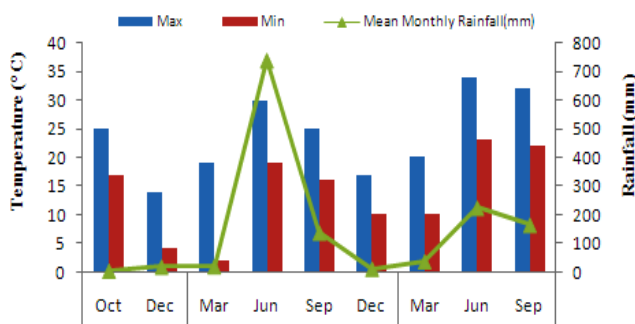


Figure I: Rainfall and temperature from Oct, 2012 to Sep, 2014 in the experimental site (Source: India Meteriological Department Site)

residual material was removed and the wet weight of the remaining litter sample was measured. The samples were then oven dried at 60 °C and reweight until constant weight achieved.

2.5 Soil study

Soil samples were randomly collected from each plot at a depth of 0-10cm and were taken to laboratory for analysis. The samples were then air dried, grounded and sieved using 2mm mesh screen. Soil pH was measured by the Orion star ion analyzer using soil:distilled water (1:2.5). Total organic carbon was analyzed using method proposed by *Walkey and Black* method¹⁵. Available and total nitrogen was estimated using *Stanford and smith*¹⁶ and Kjeldahl method, respectively. Available and total phosphorous was assessed using *Olsen* method¹⁷ and stannous chloride method¹⁴. Available potassium was determined by flame photometric method¹⁸.

Soil microbial biomass carbon, nitrogen and phosphorous (MBC, MBN and MBP) was estimated using chloroform fumigation extraction method. Final MBC, MBN and MBP were calculated by difference between fumigated and non-fumigated values with a conversion factor of 0.33 for MBC, 0.54 for MBN and 0.40 for MBP.

Soil basal respiration was determined by alkali absorption method that quantifies CO₂ evolution using moist soil sample¹⁹. Soil enzyme assay were perform in moist soil. Acid phosphates and β-glucosidase activity was measured using *Eivazi and Tabatabai* ^{20, 21}. Dehydrogenase activity was assayed by *Pepper et al.* ²² method. Protease activity was estimated using *Ladd and Butler* ²³ protocol. Fluorescein diacetate (FDA) activity was measured using method proposed by *stubberfield and Shaw*²⁴. Cellulase activity was quantified using *Schinner and Von Mersi* method²⁵.

2.6 Data analysis

The remaining dry mass for each period was calculated using the weight of litter at each sampling period and the initial weight using the formula:

$$\% \text{ RM} = (W_t/W_0) \times 100$$

Where, W_t = Weight (W_t) of litter at each sampling period (t)

W_0 = Initial mass

Decomposition rate constant (k) of each single species was calculated by the most widely used single exponential decay model from the changes of litter dry mass over time²⁶:

$$\ln (W_t/W_0) = -kt$$

Where, W_t = Amount of litter at time t

W_0 = Initial mass of litter

k = Decomposition rate constant

t = Elapsed time (year⁻¹)

The nutrient re-translocation efficiency (NRE) was computed by the equation given by *Finzi et al.* ²⁷.

$$\text{NRE \%} = \{(A - B) / A\} \times 100$$

Where, A= Nutrient in green leaves

B= Nutrient in leaf litter

Nutrient use efficiency (NUE) was calculated according to *Vitousek*²⁸.

$$\text{NUE} = \text{Litterfall mass} / \text{Nutrient content in litterfall}$$

The values were expressed in $\text{g m}^{-2} \text{yr}^{-1}$.

Data were summarized as mean \pm SD (standard deviation). Values were analyzed by Student's t test to determine the significant difference among the two stands in term of litterfall production, decomposition rate and the nutrient uptake. Soil samples were also analyzed for the significant difference among the two stands. A two-tailed ($\alpha=2$) p value less than 0.05 ($p < 0.05$) was considered to be statistically significant.

3. Results and discussion

3.1 Litter fall

Under the same climatic condition, the composition of tree species plays an important role in litter production²⁹. The two plant species found to be different in the quantity of litter biomass production. The mean annual litter fall biomass was 687.22 and 527.74 $\text{g m}^{-2} \text{y}^{-1}$ in Pine and Oak, respectively (*Figure. IIb*). A significant difference ($p < 0.05$) was found among the litter fall production by the two species. The litter fall at Pine stand is computed 23% higher than Oak stand. The maximum portion of litter fall biomass was observed in summer during April-June (*Fig. IIa*), which is characterized as dry period, and the result is in agreement with others result^{30, 1}. To have a better understanding of nutrient cycling, forest growth, successional pathways and interactions with environmental variables in forest ecosystems the evaluation of litter fall production is needed. Climate, season, substrate quality and type of vegetation vary with varying litter production^{31, 32}.

3.2 Nutrient Status

The nutrient index in green leaves and leaf litter were found to be significantly different ($p < 0.05$) in both the species. The nutrient concentration of N, P, K and Ca in green leaves and leaf litter were outlined in (*Figure. IIIa*). The variation in nutrient quality is mainly because of varied plant species differ in nutrient quality. The concentration of P, K and Ca in green leaves of Pine was recorded 0.54 mg g^{-1} , 2.15 mg g^{-1} and 12.14 mg g^{-1} , whereas the values in leaf litter were 0.32 mg g^{-1} , 1.37 mg g^{-1} and 8.69 mg g^{-1} , respectively. Similarly, the P, K and

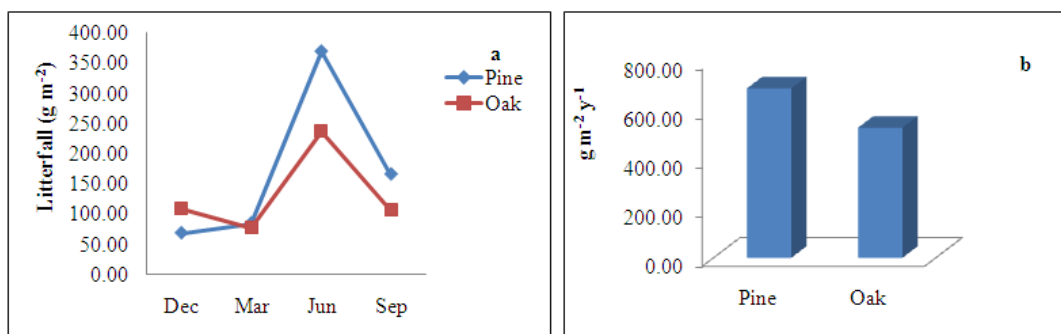


Figure II: (a) Litterfall pattern and (b) Total Litterfall in Pine and Oak stands ($\text{g m}^{-2} \text{yr}^{-1}$)

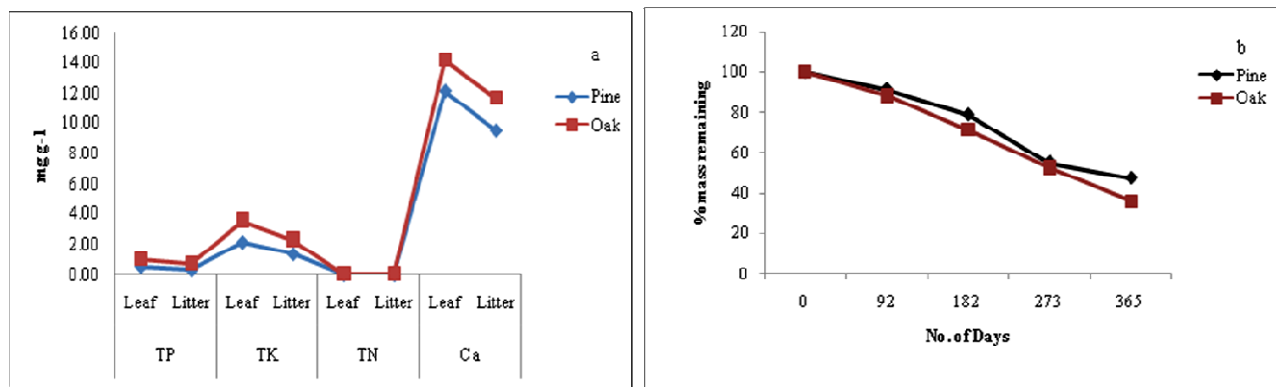


Figure III: (a) Nutrient concentration in green leaf and leaf litter of two plant species (mg g^{-1})
(b) Percentage of remaining mass at Pine and Oak stands during the study period

Ca concentration in Oak was 0.99 mg g^{-1} , 3.56 mg g^{-1} and 14.20 mg g^{-1} for green leaves and 0.58 mg g^{-1} , 2.29 mg g^{-1} and 10.36 mg g^{-1} for leaf litter, respectively. The nitrogen (%) in Pine was recorded 0.025% and 0.018% in green leaves and leaf litter. For Oak, the % nitrogen was found to be 0.030 and 0.024 in green leaves and leaf litter. Additionally, the Oak litter found to be low C/N ratio than that of Pine litter i.e. 16.78 and 21.84. This C/N ratio is a good indicator of decomposition process. A significant variation ($p < 0.05$) has been observed among all the mobile nutrients between leaf litter and green leaves. The concentration of all the nutrients was found to be low in leaf litter than the green leaves in both the plant species. The two plant species have different traits displayed species-specific variation in nutrient pattern. The plant species induced variation in chemical composition which directly affects the microbial attributes residing in that particular soil and influences the decomposition of organic matter as well as the physico-chemical status of soil^{3,33}.

3.3 Litter decomposition

The rate of decomposition is highly influence by litter quality. The litter decomposition was found to be positively correlated with the time elapsed ($r = 0.98$, $p < 0.05$). The mass remained at the end was 47.6% and 36.2% for Pine and Oak, respectively (Figure. IIIb). The calculated annual decomposition rate of Pine and Oak

litter was 0.74 and 1.01. The Pine litter has comparatively slower rate of decomposition than the Oak litter residing under same climatic condition. Pine litter with low nitrogen content and high C/N ratio results in significantly slower decomposition than Oak litter, that ultimately responsible for low weight loss % at Pine stand. The results are adjacent with other studies which have reported significantly positive correlation between C/N ratio and litter decomposition^{34,35}. The difference in the litter quality and the decomposition rate between the two stands dominated with two different plant species is mainly due to the difference among the substrate quality. Few literatures have demonstrated that stands with higher nutrient pool exhibit faster decomposition of litter³. Additionally, the efficient nutrient use efficiency is primarily described by lower nutrient concentration in litterfall²⁸. The nutrient use efficiency of the selected species specific-stands of temperate forest of Kumaun region is presented in (Fig. IVa). The nutrient i.e. N, P, K and Ca use efficiency was found to be high at Pine dominated stand. Statistical analysis revealed a significant difference ($p < 0.05$) among the nutrient use efficiency of the two species. The nutrient used efficiency is inversely related to the availability of these entire nutrients in soil. Since, it is indirectly act as an indicator of soil nutrient availability while directly related to the decomposition rate³⁶. Both the litter decomposition and the nutrient use efficiency are

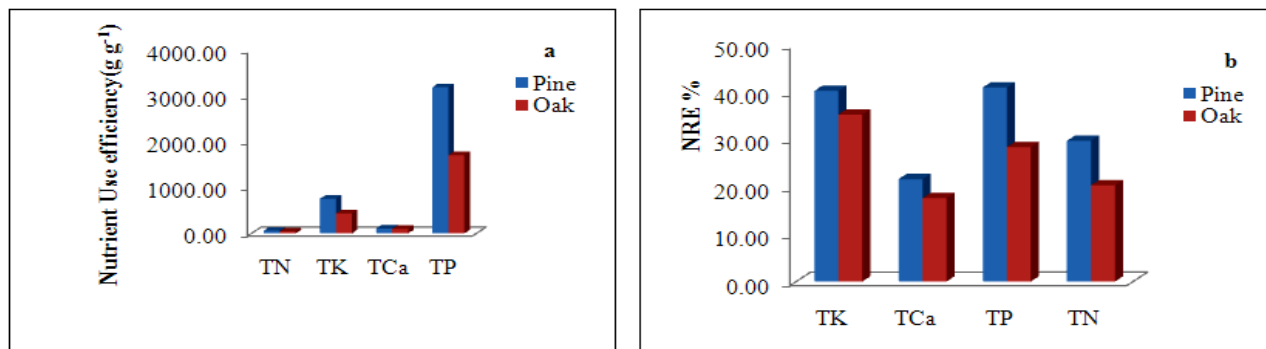


Figure IV: (a) Nutrient Use Efficiency of two different plant species and (b) Nutrient retranslocation efficiency of selected plant species

Table I: Soil physico-chemical characteristics under Pine and Oak plantation in the study area

Soil Characteristics	Pine stand	Oak stand
pH	5.687±0.015	5.867±0.031
TOC (%)	5.730±0.080	7.080±0.130
TN (%)	0.232±0.009	0.555±0.001
AN (%)	0.065±0.005	0.071±0.005
TP (mg g ⁻¹)	0.174±0.004	0.400±0.003
AP (mg g ⁻¹)	0.009±0.001	0.034±0.001
AK (mg g ⁻¹)	0.150±0.001	0.267±0.002
ACa (mg g ⁻¹)	1.067±0.003	1.770±0.003
C/N	24.72±1.26	12.75±0.28

influenced by various factors including litter composition, soil physico-chemical and biological properties. Numerous studies have suggested that recalcitrant like lignin, phenolics, tannins etc are also the chief composition of litterfall that suppresses the decomposition rate^{37, 38}. However, these parameters are not examined in the present study and needs to be investigated further.

3.4 Nutrient retranslocation and cycling

The retranslocation efficiency of nutrients for *Pinus roxburgii* and *Quercus leucotricophora* was represented in (Fig. IVb). A significant difference ($p < 0.05$) was observed in nutrient retranslocation of (N, P, K and Ca) among the two species. The retranslocation percent of N, P, K, Ca were 29.27%, 40.97%, 40.26% and 21.62% in Pine whereas in Oak the percentage of these nutrients were 20.31%, 28.37%, 35.26% and 17.61%, respectively. Several authors have also demonstrated that different tree species have differed in their nutrients retranslocation^{39, 40}. The efficient retranslocation of the essential nutrients is the most significant characteristic of the climax phase of any forest ecosystem⁴¹. The retranslocation efficiency of phosphorous and potassium

was recorded high among both the species suggested the higher remobilization of the phosphorous and potassium in the temperate forest. Furthermore, the NRE % of all the nutrients was found to be high in Pine species indicates comparatively lesser nutrient supply and soil nutrient deficient condition than that of Oak. However, the degree of retranslocation depends on its physical properties as well as their requirement by the plant. The lower NRE% in Oak species represents the lower proportion of nutrients returns to litterfall. The retranslocation with the reduction in restoration of nutrients by leaves litter, results in independence of the ecosystem with respect to soil and also a possibility of managing the available elements in timed way⁴². The difference in nutrient retranslocation is governed by several factors including plant growth, Plant species, plant age, site characteristics etc⁴³. Additionally, these responses are also dependent on the eco-physiological response of tree species with the site environment and are the strong indicator of soil fertility⁴⁴.

3.5 Soil nutrient status

The pH of both the stands was recorded acidic in nature. The detailed description of soil nutrient status of both the stands was presented in (table I, II). The nutrient status of Oak stand is found to be significantly superior to Pine stand. Significant differences ($p < 0.05$) were recorded between the two stands except AN ($p > 0.05$). The C/N ratio is comparatively high at Pine stand which is mainly related to decomposition rate and pattern of nutrient immobilization. The decrease in value reveals increase in decomposition process. Soil Microbial biomass (Carbon, nitrogen and phosphorous) and basal respiration was higher at Oak stand as compared to Pine stand. Furthermore, the Oak stand displayed higher enzymes activity than Pine stand. There is significant difference recorded among the values of all the enzymes between the two stands. The higher microbial biomass and enzymatic activities at Oak stand suggested the higher availability of nutrient at Oak stand than that of Pine stand. The plants grow within nutrient-poor soil

Table II: Soil microbial characteristics enzymatic activities under Pine and Oak plantation in the study area

Microbial Characteristics	Pine Stand	Oak Stand
MBC($\mu\text{g g}^{-1}$)	335.20 \pm 2.42	448.22 \pm 2.31
MBN($\mu\text{g g}^{-1}$)	72.80 \pm 1.51	91.38 \pm 2.18
MBP($\mu\text{g g}^{-1}$)	34.47 \pm 1.60	62.17 \pm 1.54
Basal respiration ($\mu\text{g CO}_2\text{g}^{-1}\text{ soil hr}^{-1}$)	23.22 \pm 0.21	25.54 \pm 0.21
Enzymes Assay		
Acid Phosphates	2287.70 \pm 44.64	3235 \pm 121.70
β -Glucosidase	586.34 \pm 20.52	1621.78 \pm 70.03
DHA	224.79 \pm 13.53	623.09 \pm 8.88
Protease	3702.48 \pm 39.00	4022.73 \pm 41.00
Cellulases	0.40 \pm 0.02	0.73 \pm 0.03
FDA	356.05 \pm 6.63	491.77 \pm 24.15

MB(C), MB(N), MB(P)-Microbial Biomass (Carbon), (Nitrogen), (Phosphorous)., DHA-Dehydrogenase., FDA-Fluorescein diacetate.

displayed high nutrient resorption resultant to low litter quality and decomposition rates⁴⁵ that leads to the lower magnitude of nutrient return to soil and litter. It has been observed that broad leaved tree may change the forest microenvironment in order to have faster decomposition and this was in favor of oak dominated stand where microbial activity is high^{46,35}.

4. Conclusion

The study emphasizes that the two plant species differ in litterfall plants displayed different rate of decomposition. Also, the retranslocation of nutrients influenced not only by litter decomposition but also by the soil properties as well as plant nutrient status. It has been concluded that the high enzymatic activity and microbial biomass at Oak stand signifies that the availability of nutrients was higher at Oak stand than the Pine stand. Additionally, the Pine stand was undersupplied nutrient ecosystem characterized as high nutrient use efficiency and retranslocation of nutrients, lead to the depletion in litter quality as well as decomposition rate on the floor of Pine stand. Furthermore, the nutrient status of leaves, litter and soil play equal role in the working of nutrient cycle and the survival of forest ecosystem.

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Conflict of interest

The author's declares none.

References

1. Kumar S & Tewari LM. Pattern of litter fall and litter decomposition in a *Quercus leucotrichophora* A. Camus forest in Kumaun Himalaya. *International Journal of Biodiversity and Conservation* **6** (2014), 108-114.
2. Biswas & Khan DK. Major nutrient cycling of two different tropical dry deciduous forest of West Bengal, India. *International Journal of Ecology and Environmental Sciences* **3** (2011), 77-81.
3. Swift MJ, Heal OW & Anderson JM. Decomposition in Terrestrial Ecosystems. Studies in Ecology. Volume 5, University of California Press, Berkeley, C.A., (1979).
4. Pande RR, Sharma G, Tripathi SK & Singh AK. Litterfall, litter decomposition and nutrient dynamics in a subtropical natural oak forest and managed plantation in northeastern India. *Forest Ecology and Management* **240** (2006), 96-104.
5. Freschet GT, Cornwell WK, Wardle DA, Elumeeva TG, Liu W, Jackson BG, Onipchenko VG, Soudzilovskaia NA, Tao J & Cornelissen JHC. Linking litter decomposition of above- and below-ground organs to plant-soil feedbacks worldwide. *Journal of Ecology* **101** (2013), 943-952.
6. Weltzin JF, Keller JK, Bridgham SD, Paster J, Allen BP & Chen J. Litter controls plant community composition in a northern fen. *Oikos* **110** (2005), 537-546.
7. Wang Q, Wang S & Huang Y. Comparisons of litterfall, litter decomposition and nutrient return in a monoculture *Cunninghamia lanceolata* and a mixed stand in southern China. *Forest Ecology and Management* **255** (2008), 1210-1218.
8. Fioretto A, Papa S & Fuggi A. Litterfall and litter decomposition in a low Mediterranean shrub land. *Biology and Fertility of Soil* **39** (2003), 37-44.
9. Parzych A, Astel A & Trojanowski J. Fluxes of biogenic substances in precipitation and throughfall in woodland ecosystems of the Slowinski National Park. *Archives of Environmental Protection* **34** (2008), 13-24.
10. Xu X, Enoki T, Hirata E & Tokashiki Y. Pattern and chemical composition of fine litterfall in a subtropical forest in northern Okinawa Island, Japan. *Basic and Applied Ecology* **4** (2003), 229-237.
11. Singh JS. Sustainable development of the Indian Himalayan region: Linking ecological and economic concerns. *Current Science* **90** (2006), 784-788.
12. Singh SP & Singh JS. Structure and function of the Central Himalayan oak forests. *Proceeding of Indian Academy of Science (Plant Science)* **96** (1986), 159-189.

13. Black CA. Methods of Soil Analysis. Agron. Monogr. 9, ASA, Madison, WI, USA, (1965).
14. Sparling GP, Whale KW & Ramsay AJ. Quantifying the contribution from the soil microbial biomass to the extractable P levels of fresh and air dried soils. *Australian Journal of Soil Research* **23** (1985), 613-621.
15. Walkley A & Black IA. An examination of the Degtjareff method for determining soil organic matter and a proposed modification of chromic acid titration method. *Soil Science* **37** (1934), 29-38.
16. Stanford G & Smith SJ. Oxidative release of potentially mineralizable soil nitrogen by acid permanganate extraction. *Soil Science* **126** (1978), 210-218.
17. Olsen S, Watanabe FS & Bowman RA. Evaluation of fertilizer phosphate residues by plant uptake and extractable phosphorus. *Soil Science Society of American Journal* **47** (1983), 952-958.
18. Jackson ML. Soil Chemical analysis. Englewood Cliffs, New Jersey: Prantice Hall Inc. 1958.
19. Li FD, Yu ZN & He SJ. Experimental Technique of Agricultural Microbiology, 1996, Pp. 122-123.
20. Eivazi F & Tabatabai MA. Phosphatases in soils. *Soil Biology & Biochemistry* **9** (1977), 167-172.
21. Eivazi F & Tabatabai MA. Glucosidases and galactosidases in soils. *Soil Biology & Biochemistry* **20** (1988), 601-606.
22. Pepper IL, Gerba CP & Brendecke JW. Environ. Microbiology: A Laboratory Manual. Academic Press, New York, (1995)
23. Ladd JN & Butler JHA. Short-term assays of soil proteolytic enzyme activities using proteins and dipeptide derivatives as substrates. *Soil Biology & Biochemistry* **4** (1972), 19-30.
24. Stubberfield LCF & Shaw PJA. A comparison of tetrazolium reduction and FDA hydrolysis with other measurements of microbial activity. *Journal of Microbiology Methods* **12** (1990), 151-162.
25. Schinner F & Von Mersi W. Xylanase, CM-cellulase and invertase activity in soil: an improved method. *Soil Biology & Biochemistry* **22** (1990), 511-515.
26. Olson JS. Energy storage and the balance of producers and decomposers in ecological systems. *Ecology* **44** (1963), 322-331.
27. Finzi AC, Allen AS, DeLucia EH, Ellsworth DS & Schlesinger WH. Forest litter production, chemistry and decomposition of following two years of free-air CO₂ enrichment. *Ecology* **82** (2001), 470-484.
28. Vitousek PM. Litterfall, nutrient cycling, and nutrient limitation in tropical forests, *Ecology* **6** (1984), 285-298.
29. Sundarapandian SM & Swamy PS. Litter production and leaves-litter decomposition of selected tree species in tropical forests at Kodayar in the Western Ghats, India. *Forest Ecology and Management* **123** (1999), 231-244.
30. Rawat N, Nautiyal BP & Nautiyal MC. Litter production pattern and nutrients discharge from decomposing litter in a Himalayan alpine ecosystem. *New York Science Journal* **2** (2009), 54-67.
31. Vitousek PM, Turner DR, Parton WJ & Sanford RL. Litter decomposition on the Mauna Loa environmental matrix, Hawaii: Patterns, mechanisms and Models. *Ecology* **75** (1994), 418-429.
32. Zhou G, Guan L, Wei X, Zhang D, Zhang Q & Yan J. Litterfall production along successional and altitudinal gradients of subtropical monsoon evergreen broadleaved forests in Guangdong, China. *Plant Ecology* **188** (2007), 77-89.
33. Menyailo OV, Hungate BA & Zech W. Tree species mediated soil chemical changes in a Siberian artificial afforestation experiment - Tree species and soil chemistry. *Plant and Soil* **242** (2002), 171-182.
34. Rostamabadi A, Tabari M, Salehi A, Sayad E & Salehi A. Comparison of nutrition, nutrient return and nutrient retranslocation between stands of *Alnus subcordata* and *Taxodium distichum* in Tashbandan, Amol (Mazandaran). *Journal of Wood Forest Science and Technology* **17** (2010), 65-78.
35. Tripathi SK, Sumida A, Shibata H, Ono K, Uemura S, Kodama Y & Hara T. Leaves litterfall and decomposition of different above-and belowground parts of birch (*Betula ermanii*) trees and dwarf bamboo (*Sasa kurilensis*) shrubs in a young secondary forest in Northern Japan. *Biology and Fertility of Soils* **43** (2006), 237-246.
36. Baligar VC, Fageria NK & He ZL. Nutrient use efficiency in plants *Communications in Soil Science and Plant Analysis*. **32** (2001), 921-950.
37. Prescott CE, Hope GD & Blevins LL. Effects of gap size on litter decomposition and soil nitrate concentrations in a high-elevation spruce-fir forest. *Canadian Journal of Forest Research* **33** (2004), 2210-2220.
38. Tateno R, Tokuchi N, Yamanaka N, Du S, Otsuki K, Shimamura T, Xue Z, Wang S & Hou Q. Comparison of litterfall production and leaves litter decomposition between an exotic black locust plantation and an indigenous oak forest near Yanan on the Loess Plateau, China. *Forest Ecology and Management* **241** (2007), 84-90.
39. Rouhi-Moghaddam E, Hosseini SM, Ebrahimi E,

- Tabari M & Rahmani A. Comparison of growth, nutrition and soil properties of pure stands of *Quercus castanifolia* and mixed with *Zelkova carpinifolia* in the Hyrcanian forests of Iran. *Forest Ecology and Management* **255** (2008), 1149-1160.
40. Hashemi SF, Hojati SM, Hosseini SM & Jalilvand H. Comparison of nutrient elements and elements retranslocation of *Acer velutinum*, *Zelkova carpinifolia* and *Pinus brutia* in Darabkola-Mazandaran. *Iranian Journal of Forest* **4** (2012), 175-185.
41. Staff H & Berg B. Plant litter input to soil in terrestrial nitrogen cycles. F. E. Clark and T. Roswall (eds). *Eco. Bull. (Stockholen)*, **33** (1981), 147-162.
42. Mellilo JM. "Nitrogen cycling in deciduous forests". In Clark FE, Roswall T (eds) *Terrestrial Nitrogen Cycling*, Ecological Bulletin (Stockholm). **33** (1981), 405-426.
43. Sharma JC & Sharma Y. Nutrient cycling in forest ecosystems - a review. *Agriculture Review* **25** (2004), 157-172.
44. Wright I & Westoby M. Nutrient concentration, resorption and lifespan: leaves traits of Australian *Sclerophyll* species. *Functional Ecology* **17** (2003), 10-19.
45. Rejmankova E & Sirova D. Wetland macrophyte decomposition under different nutrient conditions: relationships between decomposition rate, enzyme activities and microbial biomass. *Soil Biology & Biochemistry* **39** (2007), 525-538.
46. Singh KP, Singh PK & Tripathi SK. Litterall, litter decomposition and nutrient release patterns in four native tree species raised on coal mine spoil at Singrauli, India. *Biology and Fertility of Soils* **29** (1999), 371-378.